

Zooplankton Monitoring 2013-2015 Summary

Zooplankton were sampled monthly from 1974 through 2015 at 22 stations in the upper SFE, extending from eastern San Pablo Bay through the Delta (www.dfg.ca.gov/delta/data/zooplankton/stations.asp). At each station, 3 gear types were used to collect zooplankton of various sizes: 1) a pump for sampling smaller zooplankton, including rotifers and copepods of the genus *Limnoithona*; 2) a modified Clarke-Bumpus (CB) net for sampling mid-sized zooplankton, including cladocerans and most copepods (net mesh 160 micron openings); and 3) a mysid net for sampling mysid shrimp (net mesh 505 micron openings). Abundance indices were calculated using data from the gear type that most effectively captures each organism and are reported as the mean number of each organism per cubic meter of water sampled (catch-per-unit effort, CPUE). Here, annual abundance indices are presented from 1974 through 2015 for the most common copepods, cladocerans, rotifers, and mysids. Annual mean abundance was calculated for selected taxa as the mean March through November CPUE, as these were the only months consistently sampled throughout the entire study period. Sixteen stations were used to calculate abundance indices, including 14 fixed stations sampled consistently since 1974 and 2 non-fixed stations sampled where bottom specific conductance was 2 and 6 mS/cm (approximate salinity of 1 and 3 psu).

Copepods

When monitoring began in the 1970s, *Eurytemora affinis* and *Acartia* spp. were the most abundant copepods in the upper San Francisco Estuary (Figure 1). *E. affinis* is a calanoid copepod introduced to the SFE long before monitoring started. It was a major food source for larval and juvenile fishes of many species and also adult planktivores, such as Delta Smelt and Threadfin Shad. However, *E. affinis* annual abundance has since declined (Figure 1), as new species have been introduced and become established in the estuary. One of the first introductions was *Sinocalanus doerrii*, a freshwater calanoid copepod, initially recorded by this study in late 1978 (Orsi et al. 1983). By summer 1979, *S. doerrii* abundance surpassed *E. affinis* summer abundance, and *S. doerrii* became the most abundant calanoid copepod in the upper estuary in most years from 1979 through 1984 (Figure 1).

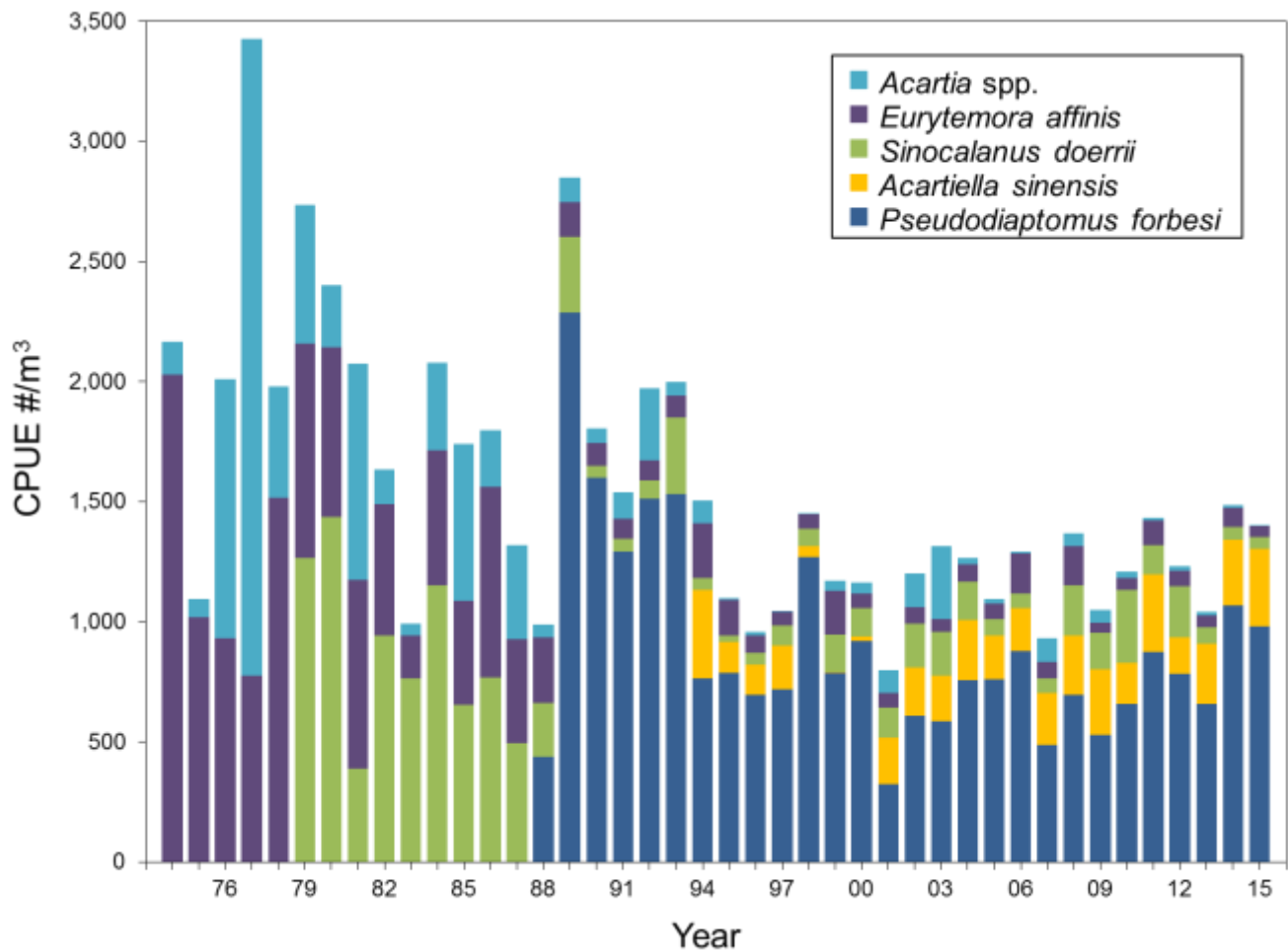


Figure 1. Annual March-November mean calanoid copepod CPUE (catch-per-unit-effort in number m^{-3}) from the Zooplankton Study's Clarke-Bumpus net for the most abundant calanoid copepods in the upper San Francisco Estuary from 1974-2015.

In 1979, the cyclopoid copepod *Limnoithona sinensis* was introduced (Ferrari and Orsi 1984). Smaller than the calanoid copepods that inhabited the estuary, *L. sinensis* (originally recorded as *Limnoithona* spp.) wasn't retained well by the CB net, but abundance estimates from the pump samples quickly became comparable to *E. affinis* and *S. doerrii* abundance estimates from the CB samples (Figure 2).

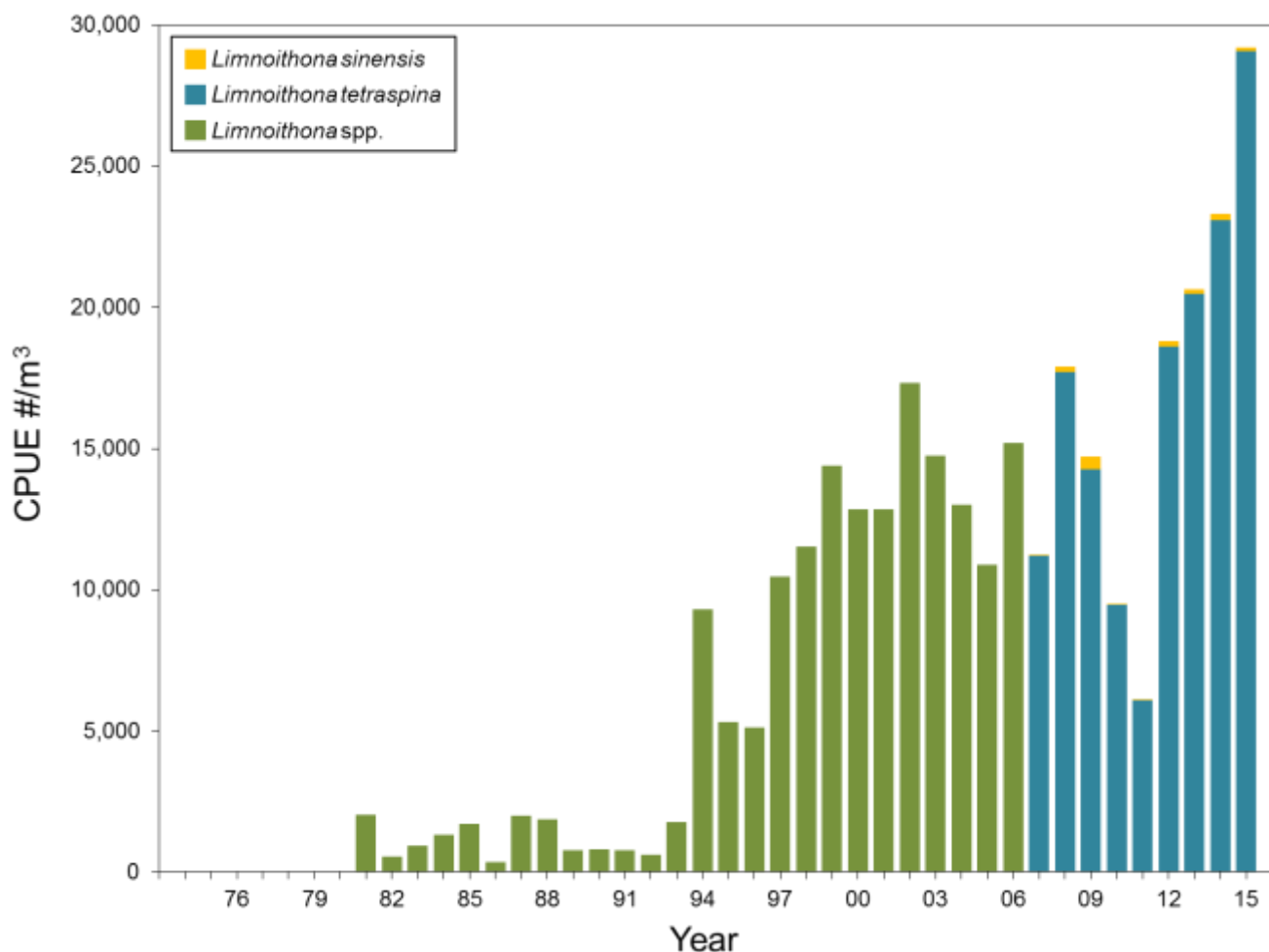


Figure 2. Annual March-November mean cyclopoid copepod CPUE (catch-per-unit-effort in number m^{-3}) from the Zooplankton Study's pump samples for the most abundant cyclopoid copepods in the upper San Francisco Estuary from 1974-2015. *Limnoithona sinensis* was originally recorded as *Limnoithona spp.*, then in 1993 *Limnoithona tetraspina* was introduced and mostly supplanted *L. sinensis*, however the genus wasn't identified to species in samples until 2007.

In the late- 1980s two more introductions occurred that further changed the copepod community in the upper SFE. In 1986 the overbite clam, *Potamocorbula amurensis*, was introduced. Then in 1987 another calanoid copepod, *Pseudodiaptomus forbesi*, was also introduced (Orsi and Walter 1991). Both *P. amurensis* and *P. forbesi* grazed on phytoplankton and thus competed with *E. affinis* for food (Kimmerer and Orsi 1996). Additionally *P. amurensis* also grazed on copepod nauplii in the water column, thereby further reducing *E. affinis* abundance through predation (Kimmerer et al. 1994). Since its introduction, *P. forbesi* has remained the most abundant calanoid copepod in the estuary during most years (Figure 1). Several more copepods were introduced in 1993. In 1993, the calanoid copepod *Acartiella sinensis* was first recorded (Orsi and Ohtsuka 1999). By 1994, *A. sinensis* was the second most abundant calanoid copepod in the upper estuary (Figure 1).

Limnoithona tetraspina, a cyclopoid copepod, was also first recorded by this study in 1993 (Orsi and Ohtsuka 1999). It mostly supplanted the historically common and slightly larger *L. sinensis* and became the numerically dominant copepod in the upper estuary by 1994 (Figure

2). *L. tetraspina* is now more abundant than any other copepod has been in the upper estuary since monitoring began, with record high annual abundances in 2013 through 2015. Despite high densities of *L. tetraspina* in the estuary, it may not be a readily available food source for visual predators, like Delta Smelt, due to its small size and relatively motionless behavior in the water column (Bouley and Kimmerer 2006). Annual abundance of *L. sinensis* was much lower than *L. tetraspina* from 2013 through 2015, but *L. sinensis* was relatively abundant compared to other copepods in the upper estuary (Figures 1 and 2).

The most abundant calanoid copepod in the upper estuary from 2013 through 2015 was *P. forbesi*, and the second most abundant was *A. sinensis* (Figure 1). Annual abundance of *P. forbesi* and *A. sinensis* was higher during 2014 and 2015, than most previous years (Figure 1). The third and fourth most abundant calanoid copepods from 2013 through 2015 alternated between *S. doerrii* and *E. affinis*, followed by the native *Acartia* spp. which were the fifth most abundant (Figure 1). Annual abundance of *S. doerrii*, *E. affinis*, and *Acartia* spp. were lower from 2013 through 2015 than in most previous years (Figure 1).

Cladocerans

Bosmina, *Daphnia*, and *Diaphanosoma* are the most abundant cladoceran genera in the upper estuary. Combined, these native freshwater cladocerans had an overall downward trend since the early 1970s (Figure 3).

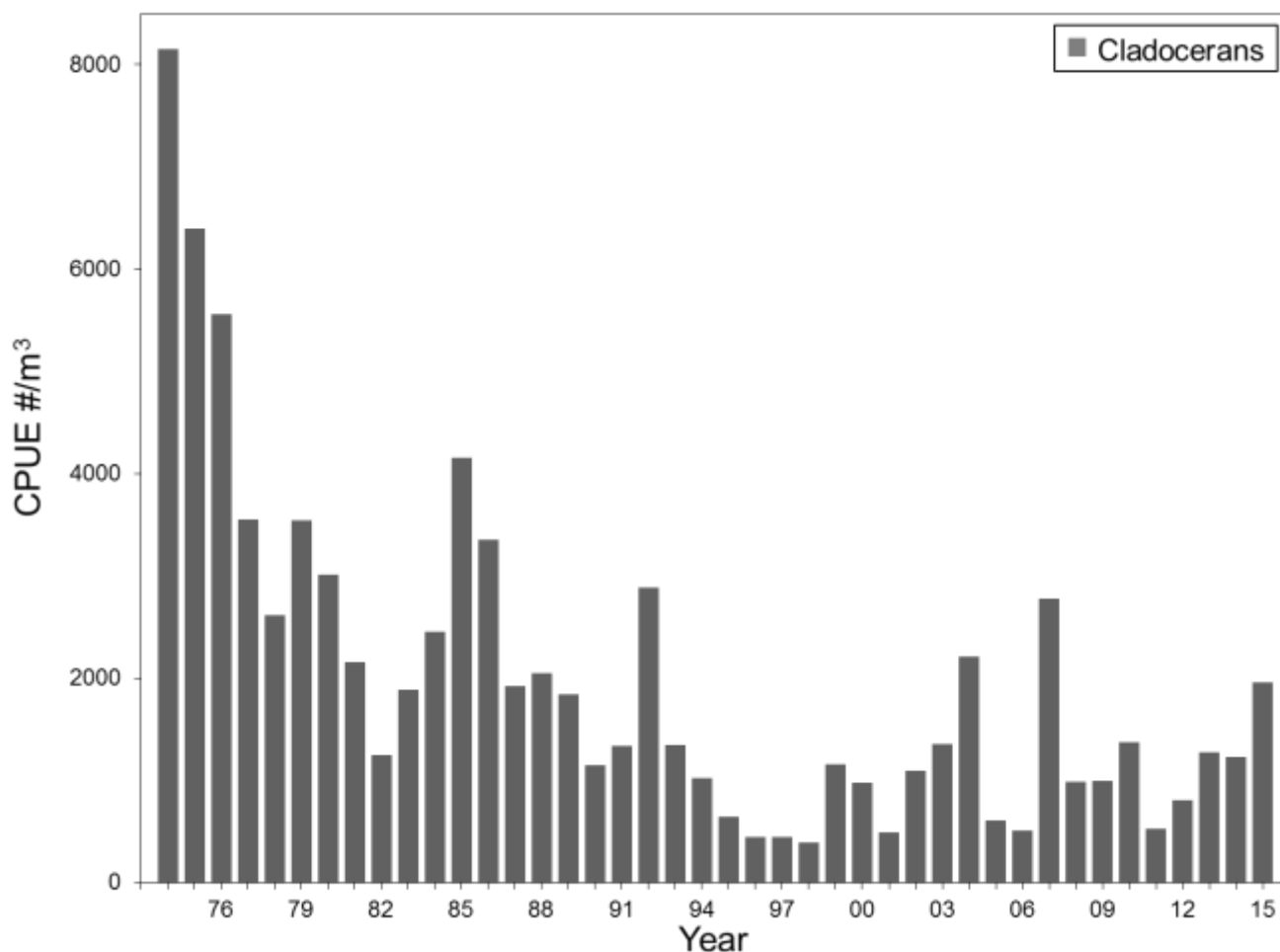


Figure 3. Annual March-November mean cladocerans CPUE (catch-per-unit-effort in number m^{-3}) from the Zooplankton Study's Clarke-Bumpus net from 1974-2015.

Rotifers

Synchaeta bicornis is a native brackish-water rotifer that has declined since the 1970s (Figure 4). In 2011 annual abundance was higher than it had been since 1985, but from 2012 through 2015 annual abundance was below average (Figure 4). Abundance of all other rotifers, without *S. bicornis*, declined from the early 1970s through the 1980s, but stabilized since the early 1990s (Figure 4). In 2011, annual abundance was higher than it had been since 1989, but declined in 2012 and despite slight increases remained below average through 2015 (Figure 4).

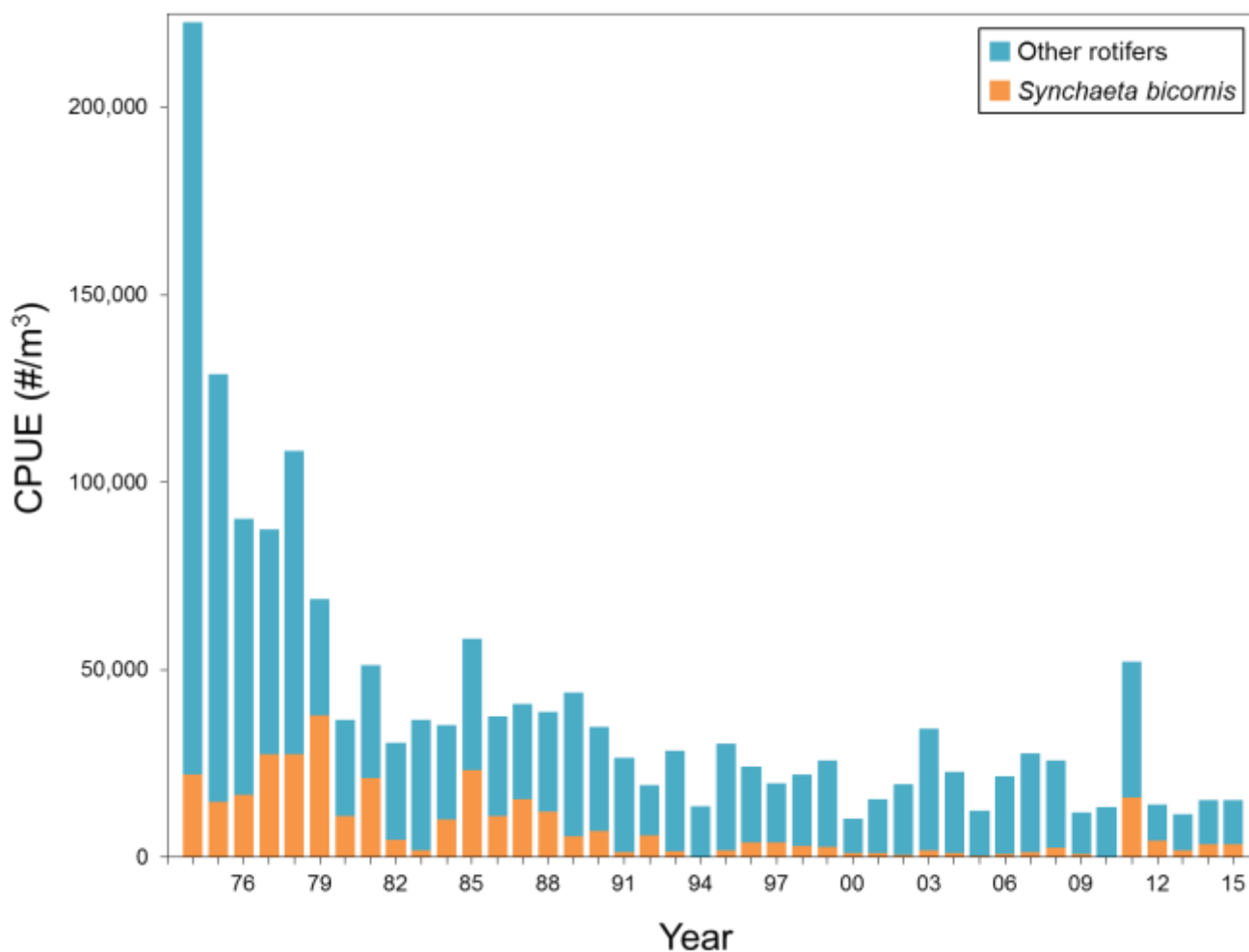


Figure 4. Annual March-November mean rotifers CPUE (catch-per-unit-effort in number m^{-3}) from the Zooplankton Study's pump samples from 1974-2015.

Mysids

Neomysis mercedis was the only mysid commonly found in the upper estuary when monitoring began in the 1970s. Similar to *E. affinis*, *N. mercedis* abundance also dropped in the early 1990s after the introduction of the overbite clam, *P. amurensis* (Table 1). This decline was

caused by competition with *P. amurensis* for phytoplankton, a shared food resource (Orsi and Mecum 1996). Shortly after *N. mercedis* abundance began declining, two newly introduced mysids were first collected by this study: *Acanthomysis aspera* was first collected in 1992, and *Hyperacanthomysis longirostris* (formerly *Acanthomysis bowmani*) was first collected in 1993 (Modlin and Orsi 1997). After its introduction, *H. longirostris* abundance increased rapidly, and by 1994 it was the most abundant mysid in the upper estuary (Table 1). In 2015, *H. longirostris* annual abundance was higher than it had been since 2002 (Table 1). *Neomysis kadiakensis* is a native brackish-water mysid that regularly appeared in mysid samples beginning in 1995, but was not abundant until recently (Table 1). *N. kadiakensis* had the second highest mean annual mysid abundance from 2013 through 2015 (Table 1). *Acanthomysis aspera*, a brackish-water mysid with historically low abundance compared to other mysids in the upper estuary, recently increased in abundance (Table 1) and from 2013 through 2015, *A. aspera* had the third highest mean annual mysid abundance (Table 1). In 2014, mean annual abundance was the highest ever recorded for *A. aspera* (Table 1). *Alienacanthomysis macropsis* is a native brackish-water mysid usually found in San Pablo Bay and Carquinez Strait that began to be consistently enumerated by this study in 1995. In 2013 through 2015, *A. macropsis* had the fourth highest mean annual abundance (Table 1). *N. mercedis*, although historically abundant before 1994, was one of the least abundant mysids in 2013 through 2015 (Table 1).

| Year | <i>Neomysis mercedis</i> | <i>Hyperacanthomysis longirostris</i> | <i>Neomysis kadiakensis</i> | <i>Acanthomysis aspera</i> | <i>Alienacanthomysis macropsis</i> |
|------|--------------------------|---------------------------------------|-----------------------------|----------------------------|------------------------------------|
| 1974 | 99.377 | | | | |
| 1975 | 76.791 | | | | |
| 1976 | 92.323 | | | | |
| 1977 | 5.420 | | | | |
| 1978 | 42.275 | | | | |
| 1979 | 35.113 | | | | |
| 1980 | 70.998 | | | | |
| 1981 | 41.946 | | | | |
| 1982 | 91.716 | | | | |
| 1983 | 17.195 | | | | |
| 1984 | 63.419 | | | | |
| 1985 | 71.184 | | | | |
| 1986 | 109.769 | | | | |
| 1987 | 26.993 | | | | |
| 1988 | 22.718 | | | | |
| 1989 | 17.448 | | | | |
| 1990 | 10.346 | | | | |
| 1991 | 16.209 | | | | |
| 1992 | 2.080 | | | | |
| 1993 | 10.254 | | | <0.001 | |
| 1994 | 0.393 | 8.100 | <0.001 | 0.108 | 0.016 |

| | | | | | |
|------|-------|--------|-------|--------|-------|
| 1995 | 0.308 | 4.046 | 0.053 | 0.001 | 0.001 |
| 1996 | 0.652 | 5.879 | 0.017 | <0.001 | 0.001 |
| 1997 | 0.202 | 14.007 | 0.142 | 0 | 0.003 |
| 1998 | 0.149 | 14.249 | 0.052 | 0.001 | 0.004 |
| 1999 | 0.182 | 17.713 | 0.040 | 0.006 | 0.005 |
| 2000 | 0.339 | 23.997 | 0.235 | 0.005 | 0.001 |
| 2001 | 0.159 | 8.990 | 0.257 | 0.017 | 0.005 |
| 2002 | 0.030 | 12.933 | 0.404 | 0.012 | 0.003 |
| 2003 | 0.022 | 9.786 | 0.230 | 0.001 | 0.014 |
| 2004 | 0.056 | 9.500 | 0.135 | 0.016 | 0.001 |
| 2005 | 0.077 | 7.870 | 0.118 | 0.003 | 0.002 |
| 2006 | 0.154 | 7.761 | 0.281 | <0.001 | 0.001 |
| 2007 | 0.009 | 3.453 | 0.200 | 0.007 | 0.010 |
| 2008 | 0.057 | 11.041 | 0.787 | 0.001 | 0.061 |
| 2009 | 0.009 | 2.597 | 0.260 | 0.001 | 0.054 |
| 2010 | 0.060 | 9.952 | 0.176 | 0.003 | 0.095 |
| 2011 | 0.163 | 2.606 | 0.235 | 0.002 | 0.039 |
| 2012 | 0.052 | 7.939 | 0.266 | 0.019 | 0.061 |
| 2013 | 0.006 | 9.034 | 0.439 | 0.032 | 0.021 |
| 2014 | 0.003 | 5.311 | 0.655 | 0.139 | 0.010 |
| 2015 | 0.002 | 12.176 | 0.886 | 0.038 | 0.004 |

Table 1. Annual March-November mean mysids CPUE (catch-per-unit-effort in number m⁻³) from the Zooplankton Study's mysid net for the most abundant mysid species in the upper San Francisco Estuary from 1974-2015.

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